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Measurement of the potential rates of dissimilatory nitrate reduction to ammonium based on 14NH4+/15NH4+ analyses via sequential conversion to N2O --Manuscript Draft--

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Dec 14, 2018

Dear Dr. Myers,

We submit the manuscript herewith entitled "Measurement of the potential rates of dissimilatory nitrate reduction to ammonium (DNRA) based on ¹⁴NH₄+/¹⁵NH₄+ analysis by sequential conversion

and detection with quadrupole GC/MS".

1) All authors have read the submitted version of the manuscript and agree to submit the work to JoVE.

2) This work has not been published before and is not being considered for publication by another

journal.

3) In this paper, we provide the detailed procedure for the measurement of potential DNRA rate and applied it to the salt marsh sediments. Briefly, the potential DNRA rate can be calculated from the ¹⁵Nlabeled ammonium (15NH₄+) accumulation rate in 15N-labeled nitrate (15NO₃-) addition incubation. Determination of non-labeled and labeled ammonium concentrations is comprised of several sequential steps, including the collection and conversion of the ammonium. These steps are basically based on previous studies, we modified several procedures. Firstly, in the step of collecting NH₄⁺, we shortened the incubation period and lowered temperature. Secondly, we made a few modifications to the method for conversion about preservation and preparation of the bacterial cells required in this procedure. These modifications allow to shorten the required time for a series of the experiments. In this method, NH₄⁺ is finally converted to N₂O and it enable to determine concentration and isotope ratio of NH₄⁺

by quadrupole GC/MS since N₂O has low atmospheric background. GC/MS is less expensive and easy to manage compared with IRMS, which has been typically used in ¹⁵NH₄⁺ studies. The representative

result demonstrated the proposal procedures effective to determine the NH₄⁺ isotope ratio and

concentration especially for samples with relatively high concentration of ammonium. Because of these, we believe that proposed method will gain broad attention from the scientific community and

make a contribution to JoVE.

We appreciate if the manuscript could be reviewed and considered for publication.

Kind regards,

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TITLE:

- 2 Measurement of the Potential Rates of Dissimilatory Nitrate Reduction to Ammonium Based on
- 3 ¹⁴NH₄+/¹⁵NH₄+ Analyses via Sequential Conversion to N₂O

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KEYWORDS:

nitrogen cycle, dissimilatory nitrate reduction to ammonium (DNRA), ¹⁵N tracer, diffusion method, persulfate oxidation, quadrupole GC/MS, salt marsh

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SUMMARY:

A series of methods to determine the potential DNRA rate based on $^{14}NH_4^+/^{15}NH_4^+$ analyses is provided in detail. NH_4^+ is converted into N_2O via several steps and analyzed using quadrupole gas chromatography—mass spectrometry.

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ABSTRACT:

The importance of understanding the fate of nitrate (NO₃⁻), which is the dominant N species transferred from terrestrial to aquatic ecosystems, has been increasing because global nitrogen loads have dramatically increased following industrialization. Dissimilatory nitrate reduction to ammonium (DNRA) and denitrification are both microbial processes that use NO₃⁻ for respiration. Compared to denitrification, quantitative determinations of the DNRA activity have been carried out only to a limited extent. This has led to an insufficient understanding of the importance of DNRA in NO₃⁻ transformations and the regulating factors of this process. The objective of this paper is to provide a detailed procedure for the measurement of the potential DNRA rate in environmental samples. In brief, the potential DNRA rate can be calculated from the ¹⁵N-labeled ammonium (¹⁵NH₄⁺) accumulation rate in ¹⁵NO₃⁻ added incubation. The determination of the ¹⁴NH₄⁺ and ¹⁵NH₄⁺ concentrations described in this paper is comprised of the following steps. First, the NH₄⁺ in the sample is extracted and trapped on an acidified glass filter as ammonium salt. Second, the trapped ammonium is eluted and oxidized to NO₃⁻ via persulfate oxidation. Third, the NO₃⁻ is converted to N₂O via an N₂O reductase deficient denitrifier. Finally, the converted N₂O is analyzed using a previously developed quadrupole gas chromatography–mass spectrometry system. We applied this method to salt marsh sediments and calculated their potential DNRA rates, demonstrating that the proposed procedures allow a simple and more rapid determination compared to previously described methods.

INTRODUCTION:

The artificial synthesis of nitrogen fertilizer and its widespread application have greatly perturbed the global nitrogen cycle. It is estimated that the transfer of reactive nitrogen from terrestrial to coastal systems has doubled since pre-industrial times¹. A significant portion of fertilizers applied to a given field is washed away from the soil to rivers or groundwater, primarily as NO_3^{-2} . This may cause environmental problems such as drinking water pollution, eutrophication, and the formation of hypoxia. NO_3^{-1} in water environments is removed from or retained in the ecosystem via biological assimilation and various microbial dissimilatory processes. Denitrification and anammox are known to be major microbial removal processes for NO_3^{-1} . Denitrification is the microbial reduction of NO_3^{-1} to gaseous N products (NO_3^{-1} , and NO_3^{-1}) coupled with the oxidation of an electron donor, such as organic substances, thereby reducing the risk of the abovementioned problems. Anammox also produces N_2 from NO_2^{-1} and NH_4^{+1} ; therefore, it removes inorganic N from an ecosystem. Conversely, DNRA works to retain N in an ecosystem; it is generally accepted that DNRA is performed primarily by fermentative bacteria or chemolithoautotrophic bacteria and that they reduce dissimilatory NO_3^{-1} to bioavailable and less mobile NH_4^{+1} .

Studies on DNRA have primarily been performed in marine or estuarine ecosystems, such as oceanic or estuarine sediments and water, salt or brackish marsh soil, and mangrove soil. Coastal or marine ecosystems are important as reservoirs for removing NO₃⁻ from terrestrial ecosystems, and in previous studies DNRA has been shown to contribute over a very wide range of NO₃⁻ removal (0–99%)³⁻¹⁸. Further, the existence of DNRA has been demonstrated in a wide range of environments including freshwater environments¹⁹, rice paddy soils²⁰, and forest soils²¹. While these studies have shown that DNRA is potentially comparable to denitrification for NO₃⁻ removal, studies measuring the DNRA activity are still very limited compared to those measuring denitrification.

The DNRA rate has been evaluated using ¹⁵N-labeling techniques in conjunction with data analysis via analytical or numerical models. One analytical solution to calculate the DNRA rate is based on the increase in the ¹⁵N enrichment of the NH₄⁺ pool after the addition of ¹⁵NO₃⁻ as a tracer. ¹⁵N-labeled NO₃⁻ is added to a sample and incubated, and the DNRA rate can then be calculated from the concentration and isotope ratio changes in NH₄⁺ before and after a certain

period of time. In this paper, a method to quantify the NH₄⁺ concentration and the isotope ratio, which are required to calculate the DNRA rate, is described in detail. Basically, the method reported here is a combination of several previously reported techniques²²⁻²⁶ with modifications added to some procedures. The method is comprised of a series of five component procedures: (1) incubation of an environmental sample with the amendment of a stable isotope tracer, ¹⁵NO₃ –, (2) extraction and recovery of NH₄⁺ using a "diffusion procedure" with modifications, (3) persulfate oxidation of NH₄⁺ in the sample, consisting of indigenous NH₄⁺ and ¹⁵NH₄⁺ derived from ¹⁵NO₃ via DNRA activity, into NO₃ and ¹⁵NO₃ , (4) subsequent microbial transformation of NO₃ and ¹⁵NO₃ to N₂O isotopomers via the modified denitrifier method, and (5) quantification of the N₂O isotopomers using gas chromatography–mass spectrometry (GC/MS). In the following section, first, the preparation for procedures (2) and (4) is described and then, subsequently, all five component procedures are described in detail.

PROTOCOL:

1. Preparation of a PTFE envelope for quantitatively capturing gaseous NH₃

1.1. Place a 60-mm piece of polytetrafluoroethylene (PTFE) tape (25 mm in width) on a small sheet of aluminum foil (approximately 300 mm x 450 mm in size, wiped with ethanol).

- 1.2. Ash a glass fiber filter (10 mm in diameter with a pore size of 2.7 μ m) at 450 °C for 4 h in a muffle furnace. Place the glass fiber filter a little above the midpoint of the longer axis of the tape (**Figure 1a**).
- 1.3. Spot 20 μL of 0.9-mol/L H₂SO₄ on the center of a GF/D filter, and immediately fold the PTFE
 tape using two tweezers: flat-ended stamp tweezers and straight-ended tweezers. The following
 steps, steps 1.4–1.7, are shown in Figure 1 and should be conducted swiftly.
- 116 1.4. Flip the PTFE tape over the GF/D filter at the dotted line shown in **Figure 1a** to form the shape shown in **Figure 1b**.
- 1.5. Seal both sides by folding and then tightly pressing the edge with the tweezers (Figure 1c).
 Do not press too hard, and do not scratch the PTFE tape.
- 122 1.6. Fold the open end with the tweezers, and then press the edge with the tweezers (**Figure 1d**).
- 1.7. Seal the open end by tightly pressing the edge with the tweezers (Figure 1e). The GF/D filter
 should not be pressed during this procedure.
 - 2. Preparing the biomass of a nitrous oxide reductase deficient denitrifier, *Pseudomonas chlororaphis* subsp. *aureofaciens* ATCC13985, for the denitrifier method
- 2.1. Streak a 20% glycerol stock of *Pseudomonas chlororaphis* subsp. *aureofaciens* ATCC13985 on
 1/4 strength tryptone soy broth (TSB) agar plates. Incubate the plates at 25 °C for 2–3 days.

- 2.2. Transfer a singles colony of *P. chlororaphis* to a small test tube containing 5 mL of autoclaved
 TSB medium and culture aerobically (without shaking) for a day at 25 °C in dark until obtaining
 maximum growth; this will be used as the preculture.
- 2.3. Transfer 3 mL of the preculture to a 1-L bottle with a silicon rubber stopper containing 1 L of freshly prepared autoclaved modified TSB supplemented with 10 mmol/L KNO₃²³. Incubate the bottle while agitating using a stirrer under dark conditions at 25 °C. After cultivating for 8 h, replace the silicon rubber stopper with a screw cap and close tightly. Continue the cultivation in anoxia overnight.
- 143 2.4. Centrifuge the culture at 18,800 x g for 15 min at 4 °C to obtain biomass pellets. 144

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- 2.5. Wash the packed biomass three times with 30 mL of Dulbecco's phosphate-buffered saline
 (D-PBS(-), pH 7.5), to completely eliminate the NO₃⁻. The conditions of the centrifugation are
 same as in Step 2.3.
- 2.6. After washing, re-suspend the packed biomass with 30 mL of D-PBS(-). Use 1 mL of the suspension to determine the cell density by measuring OD₆₀₀. Pipette 1 mL aliquots of the remaining suspension into sterile cryovials containing 0.8 mL of 45% glycerol. Preserve the prepared glycerol stocks at -80 °C until use (see section 6).

3. Elimination of oxygen, nitrite, and nitrate from the sample sediment

- 3.1. Weigh 3.0 g (wet weight) of sediment into a 20 mL glass vial and add 9.0 mL of surface water to suspend it (25% w/w slurry).
- 3.2. Seal the vial with a black butyl rubber stopper (washed with ion-exchanged water and sterilized via autoclaving) and an aluminum cap.
- 3.3. Purge the suspension and replace the headspace air with Ar (>99.99%) at 0.6 L/min for 20 min using a manifold.
- 3.4. Replace the Ar headspace gas with ultra-pure (>99.99995%) He by vacuuming for 90 s and charging He for 30 s. Repeat this procedure four times. Set the headspace gas pressure to 1.5 atm.
- 3.5. Incubate the vials at 20 °C overnight with shaking at 150 rpm under dark conditions using a constant temperature shaker to eliminate the remaining oxygen, nitrate, and nitrite in the sediment suspension and headspace gas.

4. Time course experiment for determining DNRA rate

4.1. Replace the headspace gas with fresh ultra-pure He using the same procedure as in step 3.5but without the four repetitions.

4.2. Add labeled and non-labeled substrates to each vial according to **Table 1** through the butyl rubber stopper using a gastight syringe. Purge the previously prepared substrate solutions in adequately sized glass vials using ultra-pure He with the pressure of the headspace set to 1.5 atm to avoid air contamination. Avoid the unintentional injection of any amount of air during the injection procedures.

4.3. Incubate vials at 20 °C with shaking at 150 rpm. Add the substrates to each vial according to **Table 1** and incubate for 1 h, 3 h, and 5 h. After incubation is halted, subject the sediment suspension in the vials to the following procedures: steps 4.4–4.9.

4.4. Remove the aluminum cap and butyl rubber stopper from each vial. Then, add KCl (ashed at 450 °C for 4 h) to the sediment suspension up to a final concentration of approximately 2 mol/L to ensure the recovery of NH₄⁺ from the sediment. Close the vials with a butyl rubber stopper and an aluminum closure.

4.5. Shake the sediment at 150 rpm for 1 h at 4 °C under dark conditions to extract the NH₄⁺.

4.6. Transfer the entire sediment suspension in the vial to a 50 mL plastic centrifuge tube, and centrifuge at $10,000 \times g$ for 10 min at $4 \text{ }^{\circ}\text{C}$.

4.7. Rinse the internal wall of a freshly opened 10-mL disposable syringe, and attach it to a freshly opened disposable cellulose acetate membrane filter (pore size $0.22 \mu m$, 25 mm in diameter). Then, rinse the membrane filter with 1 mL of supernatant. Place the rinsed syringe—filter unit onto a 20-mL wide-mouth polypropylene (PP) bottle.

4.8. Filter the remaining supernatant through an acetate membrane filter into the 20-mL PP bottle. Store the extracts at −20 °C until further analysis.

5. Capturing diffused NH₄⁺ in 2M H₂SO₄ absorbed to the GF/D filter in the PTFE envelope and the persulfate oxidation of NH₄⁺ to NO₃⁻

5.1. Prepare standard solutions of $^{14}NH_4Cl$ with concentration gradients of 0 µmol/L, 10 µmol/L, 40 µmol/L, 100 µmol/L, 200 µmol/L, 400 µmol/L, and 500 µmol/L. For the ^{15}N ratio analysis, fix the total concentration of NH_4^+ to 200 µmol/L and prepare isotope ratios of 100:0, 99.5:0.5, 99:1, 93:7, 90:10, 50:50, and 10:90 with pure $^{14}NH_4Cl$ and $^{15}NH_4Cl$ standard solutions.

5.2. Transfer 30 mg of MgO (ashed at 450 °C for 4 h) to a 20-mL glass vial, and place the PTFE envelope in the vial.

217 5.3. Transfer 5 mL of a sample or standard into the vial containing the MgO and the PTFE envelope, and immediately close with a gray butyl rubber stopper. Seal with an aluminum cap. If the concentration of NH_4^+ is expected to exceed 500 μ mol/L, dilute the sample to below 500 μ mol/L.

5.4. Shake the vials at 150 rpm for 3 h at 4 °C under dark conditions.

5.5. Remove the aluminum cap and the butyl rubber stopper. Take the PTFE envelope out of the vial using point-ended tweezers, thoroughly rinse the envelope and the tweezers with ion-exchanged water, wipe them with a wiping paper, and then place the envelope on a fresh wiping paper.

5.6. Open the PTFE envelope with a couple of tweezers (use of both the flat-ended and point-ended tweezers is recommended) in the exact reverse order of the folding performed in steps 1.4–1.7.

5.7. Hold the peripheral area of the GF/D filter, where the H₂SO₄ is supposed to be unabsorbed, with the flat-ended tweezers, and transfer it into an 11-mL screw cap test tube with a PTFE -lined cap. Rinse the tweezers with ion-exchanged water, and wipe them with wiping paper.

5.8. Repeat steps 5.5–5.7 for the remaining envelopes.

5.9. Add 1 mL of ion-exchanged water to each of the test tubes, close the screw cap, and maintain it without shaking for at least 30 min at room temperature to completely elute the NH₄⁺ from the GF/D filter. During this step, carry out the following step (step 5.10) in parallel.

5.10. Prepare the persulfate-oxidizing solution (POR) reagent^{25, 26}.

5.10.1. Because it cannot be stored, prepare the exact amount of POR needed for treating a single day of samples.

5.10.2. To prepare POR to treat 50 samples, pour 100 mL of ion-exchanged water into a 200-mL screw cap bottle and add 1.52 g of NaOH (nitrogen compound analysis grade), 3 g of boric acid, and 5 g of $K_2S_2O_8$ (nitrogen and phosphorus analysis grade) in that order. Immediately after adding each reagent, shake the solution until it is completely dissolved.

5.10.3. If necessary, soak the bottle in warm water to help the dissolution of the chemicals; however, attention should be paid with respect to the contamination of NO_3^- because tap water normally contains NO_3^- .

5.11. After Step 5.8, open the screw cap, add 2 mL of the POR reagent to the test tube, and close the tube tightly with a screw cap to prevent any loss or contamination during the following steps.

5.12. Stand the test tubes on a rack, wrap them in double-layered aluminum foil, and autoclave them for 1 h at 121 °C. Keep the tubes in an upright position during this step and avoid rapid changes in temperature after finishing the autoclaving.

264 6. Determining the NO₃⁻ converted from NH₄⁺ by the denitrifier method using quadrupole GC/MS

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267 6.1. Mix 100 mL of sterile 40-mmol/L phosphate buffer (pH 7.2) and 100 mL of sterile 30-mmol/L glucose aseptically (20-mmol/L phosphate and 15-mmol/L glucose; final).

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270 6.2. Add a 1/7.2 volume of glycerol stock of *P. chlororaphis* to 200 mL of the phosphate-buffered glucose solution in a 300-mL Erlenmeyer flask, and purge with an ultra-pure He (>99.995%) stream for 1 h.

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274 6.3. Dispense 2.0 mL of the denitrifier suspension into each 5-mL vial. Cap the vials with a gray butyl rubber stopper and an aluminum closure.

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277 6.4. Replace the headspace air with ultra-pure He by vacuuming for 3 min and charging the He for 1 min. Set the headspace gas positive pressure to 1.3 atm to avoid unintentional air contamination.

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6.5. Inject 1 mL of a sample or standard through the butyl rubber stopper using a 1-mL disposable syringe. Note the exact amount of the sample actually injected.

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6.6. Incubate the vials overnight at 25 °C under dark conditions.

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6.7. Inject 0.3 mL of 6-mol/LL NaOH to stop the denitrification and absorb the headspace CO_2 , which will otherwise seriously disturb the N_2O analysis by GC/MS because CO_2 and N_2O have the same molecular weight.

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6.8. Determine the amounts of $^{44}N_2O$, $^{45}N_2O$, and $^{46}N_2O$ in the headspace gas using quadrupole GC/MS with a modified injection port²⁵. The operating conditions used for the GC/MS analysis are shown in **Table 2**.

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7. Data analysis

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7.1. Derive the calibration curve for the NH_4^+ concentration from the linear relationship between the known concentration of $^{14}NH_4^+$ and the measured signal intensities of the total produced $^{44}N_2O$ + $^{45}N_2O$ + $^{46}N_2O$. Derive the calibration curve for the ^{15}N content from the linear relationship between the known atom% (*i.e.*, $^{15}N/^{14}N+^{15}N$) and the calculated atom% using a previously provided equation²⁷. Calculate the concentration of $^{15}NH_4^+$ by multiplying the total NH_4^+ by the ^{15}N ratio of NH_4^+ .

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7.2. Calculate the potential DNRA rate using equations provided elsewhere²⁸⁻³⁰.

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REPRESENTATIVE RESULTS:

The representative results presented in this paper were derived from ¹⁵N-tracing experiments of salt marsh sediments. The sampled salt marsh was newly created in the aftermath of the 2011

Great East Japan Earthquake in the Moune area of Kesen-numa city in Miyagi Prefecture, Japan. In September 2017, surface sediments (0–3 cm) were collected at two sites in the subtidal and intertidal zones. First, immediately after collection, the sediment was sieved through a 4-mm mesh to remove plant roots, shells debris, and rubble and then homogenized. The samples were stored at 4 °C until the DNRA analysis was conducted.

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The incubation procedures for the ¹⁵NO₃⁻ and the simultaneous determination of the ¹⁴NH₄⁺ and $^{15}\text{NH}_4^+$ concentrations were carried out as described in the protocol section. An increase in the ¹⁵NH₄⁺ concentration throughout the incubation period was observed for all sediments (Figure 2). We calculated the DNRA rates by dividing the accumulation rate of ¹⁵NH₄+ by the isotope ratio of the NO_3^- pool²⁹. The calculated rates were within the range of 24.8–177 nmol-N g^{-1} dry soil h^{-1} (Table 3) and were comparable to values found in previous studies. This range of obtained rates is higher than the reported values derived from similar environments including those from intertidal sediments¹⁷, salt marshes^{5, 16}, and other estuarine environments^{18, 33, 34}, as well as from eutrophic environments such as a shallow river estuary in North Carolina³¹ and the Shanghai urban river networks³². Conversely, Fernandes et al.¹³ reported higher potential DNRA rates in organically rich mangrove soils in India. In general, DNRA is thought to be favored by a high ratio of available C to electron acceptors³⁵⁻³⁷. The samples demonstrating the representative results were taken from a salt marsh newly created by an earthquake, which had originally been used as a cultivation field. This particular characteristic of the samples may contribute to the observed high DNRA rate. Consistent with this speculation, the DNRA rate in the intertidal zone, which is rich in organic compounds (data not shown) compared to the subtidal zone, was higher than that in the subtidal zone (Figure 2, Table 3).

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FIGURE & TABLE LEGENDS:

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Figure 1: Preparation of a PTFE envelope for capturing gaseous NH₃. The PTFE envelope used in the diffusion procedure is prepared by folding PTFE tape following the instructions shown in panels **(A)–(E)**. The acidified filter inside the envelope captures the gaseous NH₃. These steps should be conducted quickly. Detailed information is given in steps 1.2–1.7 in the protocol section.

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Figure 2: Change in the $^{15}NH_4^+$ concentration via the anaerobic incubation of sediments. The ^{15}N tracer incubations of the sediment samples were conducted in duplicate. The concentration of $^{15}N-NH_4^+$ is shown in nmol per dry weight of sediment.

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Figure 3: An example of calibration curve of low concentration N_2O . Peak area of N_2O was obtained by sum of the peak area of m/z 44, m/z 45, and m/z 46. Configurations for GC/MS analysis is shown in Table 2.

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Table 1: Combinations of substrates amended to vials retaining approximately 11 mL of the sediment suspension. Samples were prepared in duplicate and subjected to further analyses after 1 h, 3 h, and 5 h of incubation.

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Table 2: Conditions for the GC/MS analysis.

Table 3: Potential DNRA rates of the tested intertidal and subtidal sediments.

DISCUSSION:

The concentration and isotope ratio of NH_4^+ for the DNRA analysis was quantified using several methods. The concentrations and isotope ratios of NH_4^+ are generally measured separately. The NH_4^+ concentration is typically measured using colorimetric methods including an autoanalyzer^{4,10,15-17}. The isotope ratio measurement has wide variations depending on its method of NH_4^+ conversion, trapping, and instrumentation for the analysis. Typical methods include the following:

- (1) The NH_4^+ in the sample is converted to NH_3 via the addition of MgO or NaOH. After moving from the liquid phase to the gas phase, the NH_3 is trapped on an acidified glass filter or in an acid solution. After drying, the filter is combusted and analyzed as N_2 using an elemental analyzer/isotope ratio mass spectrometer (EA-IRMS)^{11,18,22,38}. Alternatively, the captured NH_4^+ in an acid solution is collected on an adsorbent (e.g., zeolite) and is combusted and analyzed via EA-IRMS^{10,12}.
- (2) NH_4^+ is oxidized to N_2 via hypobromite oxidation. The isotopic composition of the evolved N_2 is measured via IRMS^{4,14,39} or membrane-introduction mass spectrometry (MIMS)^{16,17}.
- (3) The NH_4^+ concentration is determined using high-performance liquid chromatography without any conversion. Because the equilibrium of NH_4^+ and NH_3 is slightly different for $^{15}NH_4^+$ and $^{14}NH_4^+$ near the pH of pK_a, the $^{14}NH_4^+$ and $^{15}NH_4^+$ concentrations can be determined based on a small shift in the retention time^{6,19,40}.

The method described in this paper is basically the same as approach (1) listed above, i.e., the step recovering NH₄⁺ as NH₃ on a glass filter under alkaline conditions; however, it is different with respect to the following sequential NH₄⁺ conversion steps. These conversion steps are based on previous studies with added modifications to shorten the required time for a series of experiments. First, we made a few modifications to the original denitrifier method. After the biomass of *P. chlororaphis* is prepared as previously described^{23,24}, we grow bacteria and preserve the concentrated cell suspension as a glycerol stock. This dense cell suspension can be directly used for the isotope analysis by mixing it with a buffer solution because the denitrifying activity has already been sufficiently induced. Even though further investigation is required, this modification may improve the reproducibility of the analysis because the presented method enables the mass-cultivation of denitrifier cells, which are directly available for analyses. We also modified the composition of the solution for suspending the bacterial cells, from a medium based on TSB to a phosphate-buffered glucose solution, to exclude the unnecessary components such as peptone in the original medium. This modification may reduce contamination by blank N because the phosphate-buffered glucose solution does not contain N, unlike the original medium used in previous studies; this should be tested via further analyses. In the step collecting NH₄⁺ via the diffusion method, we shortened the incubation period and lowered the temperature to minimize the breakdown of organic N and any unfavorable conversion or loss of NH₄⁺. The validity of this modification was checked using the linearity of the calibration curve. We also checked that the modified temperature and incubation period did not affect the recovery of NH_4^+ (data not shown).

Another advantage of this method is that NH_4^+ is ultimately converted to N_2O , which has a low atmospheric background and can be measured using quadrupole GC/MS, which is less expensive and easier to manage than IRMS. Under the condition shown in table 2, CO_2 (m/z 44) and N_2O (m/z 44) are completely separated by GC; the retention time of these gases are 1.15 and 1.07 min, respectively. Since atmospheric concentration of N_2O is on the order of ppb, N_2O can be measured with negligible air interference even in low concentrations. The calibration curve of N_2O passes almost through the origin, demonstrating the influence on the concentration of N_2O due to atmospheric contamination is very limited (**Figure 3**). This method also has the advantage that it can quantify the $^{14}NH_4^+$ and $^{15}NH_4^+$ concentrations together; canonical methods, except approach (3) listed above, require individual analyses for the concentration and the isotope ratio.

Overall, the limits of detection and quantification for NH_4^+ using this method were approximately 0.03 µmol and 0.09 µmol, respectively, and these values are equivalent to 6 µmol/L and 18 µmol/L, if 5 mL of the sample solution (i.e., the sediment extract in this case) is used for the diffusion procedure as described in this paper. Even though using the colorimetric method is recommended to determine the NH_4^+ concentration of samples that have low NH_4^+ , the proposed method effectively determines the NH_4^+ isotope ratio and the concentration in samples with relatively high concentrations of ammonium.

ACKNOWLEDGMENTS:

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DISCLOSURES:

The authors have nothing to disclose.

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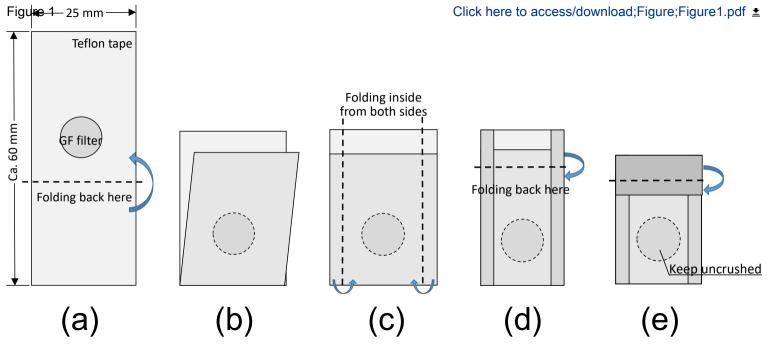
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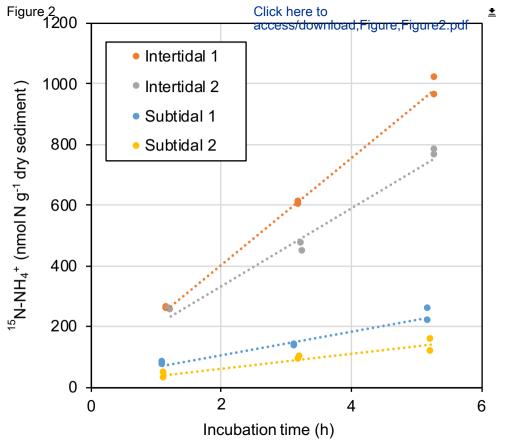
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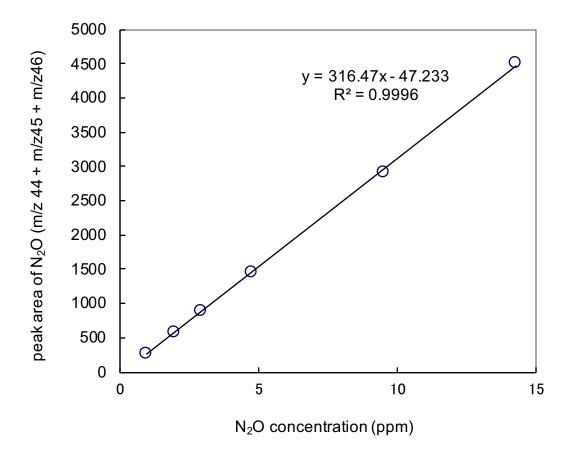
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-	¹⁵ N-labeled and non-labeled substrates added to each vial	
_	100mM	100mM
	NH ₄ CI	K ¹⁵ NO ₃
Volume (µL) of 100 mM stock solution added to each vial	24	60
Final concentration (µmol/L)*	>230 [§]	570

^{*}shown values are calculated by assuming that water content of the sediment is 50% §depending on the background ammonium concentration

Equipment

column

quadrupole GCMS Shimadzu GCMS-QP2010 Ultra

CP-PoraBONDQ 25m; φ 0.32mm; film

thickness, 5µm

Analytical conditions

column temp 40 °C injection port temp 100 °C

carrier gas stream Total flow rate, 47.1 mL·min⁻¹

flow rate in column, 2.10 mL·min⁻¹

sprit ratio 20 detection voltage 1.5 kv

Sensitivity of N₂O

lower limit of detection (LOD)* 1.42 pmol lower limit of quantification (LOQ)* 4.58 pmol

*LOD and LOQ were determined by a linear relationship among a serial dilution of N_2O (0.97, 1.94, 2.91, 4.75, 9.50, 14.3 ppm) in He, corresponding responses in peak area, and S/N ratio. LOD and LOQ were calculated as concentrations equivalent to S/N=3 and S/N=10, respectively.

sediment D		DNRA rate	enrichment of NO ₃
		nmol-N g ⁻¹ hr ⁻¹	atom%
	Intertidal 1	177	99.9
	Intertidal 2	129	99.0
	Subtidal 1	39.3	99.9
	Subtidal 2	24.8	99.0

^{*}same as atom% of added KNO₃; complete elimination and no nitrification under used incubation conditions was tested previously.

Name of Material/	Equipment	Company

 $^{15} \text{N-KNO}_3 \hspace{1cm} \text{SHOKO SCIENCE} \\ ^{15} \text{N-NH}_4 \text{Cl} \hspace{1cm} \text{SHOKO SCIENCE} \\ 20 \text{ mL PP bottle} \hspace{1cm} \text{SANPLATEC} \\ \text{Aluminum cap} \hspace{1cm} \text{Maruemu} \\ \text{Boric acid} \hspace{1cm} \text{Wako} \\ \text{Centrifuge} \hspace{1cm} \text{HITACHI} \\ \end{array}$

Deoxygenized Gas Pressure & RepliSANSIN INDUSTRIAL

Disposable cellulose acetate memb ADVANTEC
Disposable syringe Termo
Disposable syringe Termo

Dulbecco's Phosphate Buffered Sal NISSUI PHARMACEL Gastight syringe VICI Valco Instrume

GC/MS shimadzu GF/D Whatman Glass vial Maruemu Gray butyl rubber stopper Maruemu Wako H_2SO_4 $K_2S_2O_8$ Wako KCl Wako Wako KNO₃ NaOH Wako NH₄CI Wako Plastic centrifuge tube **ASONE**

Pseudomonas chlororaphis subsp. ¿American Type Cult PTFE sealing tape Sigma-Aldrich

Reciprocating shaker TAITEC
Screw-cap test tube IWAKI
PTFE-lined cap for test tube IWAKI

Tryptic Soy Broth Difco Laboratories

Catalog Number

Comments/Description

N15-0197

N15-0034

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021-02195

Himac CR21G II

IP-12

25CS020AS

SS-10SZ SS-01T

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84-0252

84-0262 211825 Wide-mouth

No. 20, with hole

Pore size $0.22 \mu m$, 25 mm in diamete

10 mL

1 mL

Series A-2, 100 µL

10 mm in diameter

20 mL

No.20-S

Guaranteed Reagent

Nitrogen and Phosphorus analysis gra

Guaranteed Reagent

Guaranteed Reagent

Nitrogen compounds analysis grade

Guaranteed Reagent

50 mL, VIO-50BN

Freeze-dried, the type strain of Pseuc

25 mm in width

NR-10

11 mL

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lomonas aureofaciens

Title of Article:



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Protocol:

1. Please ensure that all text in the protocol section is written in the imperative tense as if telling

someone how to do the technique (e.g., "Do this," "Ensure that," etc.). The actions should be described in the imperative tense in complete sentences wherever possible.

ANSWER: As the editor pointed out, we have made some sentence reconstructions to be in the imperative tense from a passive voice.

2. There is a 10 page limit for the Protocol, but there is a 2.75 page limit for filmable content.

Please highlight 2.75 pages or less of the Protocol (including headers and spacing) that identifies the essential steps of the protocol for the video, i.e., the steps that should be visualized to tell the most cohesive story of the Protocol. Remember that non-highlighted Protocol steps will remain in the manuscript, and therefore will still be available to the reader.

ANSWER: We apologize for our carelessness. We have highlighted the essential steps for video filming.

3. For each protocol step, please ensure you answer the "how" question, i.e., how is the step performed? Alternatively, add references to published material specifying how to perform the protocol action. If revisions cause a step to have more than 2-3 actions and 4 sentences per step, please split into separate steps or substeps.

ANSWER: We have revised the protocol section to meet the journal requirement.

Specific Protocol steps:

1. 1.2: Please explain the ashing procedure further or include a reference.

ANSWER: We have added a sentence for further explanation for ashing process to 1.1.2..

2. 2.1: Where do P. chlororaphis colonies come from?

ANSWER: We have added a description about incubation of *P. chlororaphis* on agar plate as step 2.2.1. in the protocol section.

Figures:

1. Figure 2: Please use 'h', not 'hr'.

ANSWER: Addressed.

Discussion:

1. Please discuss critical steps in the protocol as well as limitations of the protocol in the Discussion section.

ANSWER: We discussed our novelty and validity in line 375-401(changes accepted version). This

part includes the description of our procedures modified from previous methods. We mentioned the limitation of our protocol in line 403-409(changes accepted version).

References:

1. Please ensure references have a consistent format. Please do not abbreviate journal titles.

ANSWER: We have checked reference list and modified to meet a consistent format.

Table of Materials:

1. Please ensure the Table of Materials has information on all materials and equipment used, especially those mentioned in the Protocol.

ANSWER: We checked and modified the list of Table of Materials so that all commercial products/equipments are sufficiently included in the list.

Reviewer #1:

Manuscript Summary:

Excellent idea but need more data from GC MS analysis

ANSWER: We appreciate for your positive comment and advise.

Major Concerns:

One major concern is I need to see the GC MS data. You have mentioned in line 382 that N2O can be measured in GCMS without any atmospheric interference. Can you back that up with the MS data?

ANSWER: Under the condition shown in table 2, CO_2 and N_2O are completely separated by GC; the retention time of these gases are 1.15 and 1.07 min. respectively. In order to demonstrate that N_2O can be measured with negligible air interference even in low concentrations, we provide an example of calibration curve of low concentration N_2O (Fig. 1, in this document). Since atmospheric concentration of N_2O is on the order of ppb, the influence on the quantitative determination of N_2O due to atmospheric contamination associated with measurement is very limited. The calibration curve passes almost through the origin, demonstrating the influence on the concentration of N_2O due to atmospheric contamination is negligible (Fig.1).

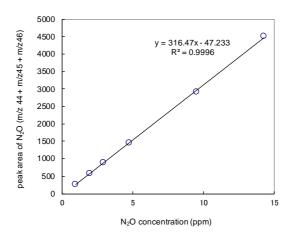


Fig. 1 The relation between N_2O concentration (ppm) and GC/MS peak area. Peak area of N_2O was obtained by sum of the peak area of m/z44, m/z45, and m/z 46.

Reviewer #2:

Manuscript Summary:

The method described will be useful to the community interested in DNRA. I am sure it will be of value to the community of researchers interested in comprehensively addressing the entire N-cycle in various ecosystems. Of particular interest is the potential to increase sensitivity by quantifying N2O instead of ammonium directly. Although this approach requires a series of steps the details are provided very clearly.

ANSWER: We appreciate for your positive comments for significance of our paper.

The Major Concerns:

The most serious issue with this manuscript is the poor quality of the english writing. Frankly it made it difficult to read at times. The detailed methods, however, were sufficient written to make them clear. The remaining sections require significant editing.

ANSWER: We apologize the poor quality of our English. We have used English language editing service and we have made modifications to the manuscript in terms of language and grammar. We have also made some sentence reconstructions and word choice changes for accuracy and enhanced clarity. We believe that these revisions improve readability of our manuscript.